# Report on Alberta Grizzly Bear Assessment of Allocation 

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## Executive Summary

A review of available data was undertaken to assist in management of the harvest of grizzly bears in Alberta. A small team of experts was tasked with this review under the direction of the Minister of Sustainable Resource Development. The team's goals included the review of the current model being used to allocate hunting licenses, an analysis of the available mortality data held by the department, and the identification of data needs to assist in grizzly bear management efforts.

The review found that the model currently used in predicting population size and thus the allocation of licenses is incomplete and will continue to predict exponential growth rates when this is not biologically possible. Solutions and modifications to this model are suggested that will improve the allocation of hunting licenses to prevent possible over harvest.

The mortality data used in this report is not as strong as it could be given the current registration process. Minor but important changes are possible to strengthen this data for future management needs and are outlined in this report. The ongoing maintenance and evaluation of a comprehensive database is seen as a key component of all current and future grizzly bear management efforts.

Our review of the available mortality data suggests that there are three Bear Management Areas (BMA' s) that exhibit a declining trend in the age of females in total mortalities. (units 2B, 4B and 4C). This represents a potentially serious management concern and steps to reduce overall mortality rates in these BMA's are necessary. In other BMA's a conservative legal harvest can be maintained if additional inventory work is undertaken and current habitat conditions are revisited to update our allocation formula. Further the review revealed that female bears continue to form a major component of overall total mortalities and measures to reduce this are necessary.

Problem bear management actions continue across the province, however we have identified three key areas in which ongoing and regular problems persist. Efforts should be made to work with stakeholders in these areas to reduce problem bear occurrences and department staff directed to ensure that conservation priorities are a key component of problem bear management.

Other issues concerning grizzly bear hunting are discussed and suggestions are made on approaches to resolve aspects that result in an overall increase in total human-caused mortality in the province.

We see opportunities to rectify current deficiencies in the data necessary to justify hunting license allocations. These include updating recent inventory data on this species, completing a comprehensive mortality/relocation database, and tracking landscape changes that affect the habitat carrying capacity for each BMA. In order to improve our grizzly bear management capability, additional resources are required in these areas.

## List of Tables

1. Number of grizzly bears estimated to occur in Alberta BMA's (1988-2002).

1b. Carrying capacities for BMA's suggested in 1990 Management Plan.
2. Example of calculation of grizzly bear population estimates for the period 19882001 using suggested modified formula.
3. Known grizzly bear mortality rates in Alberta (1972-2002).
4. Grizzly Bear Management Areas Combined for Analysis.
5. Frequencies of samples aged and not aged for removal and mortality data by pooled years and BMA's.
6. Mortalities and frequencies of aged bears for selected WMU's.
7. Summary of harvest and licenses allocated by GBMA's 1998-2002.

## List of Figures

1. Bear Management Areas (BMA's) in Alberta.

1b. Recalculation of grizzly bear population estimates using revised formula.
2. Known grizzly bear mortalities by year 1997-2002.
3. Known grizzly bear mortalities by BMA 1997-2002.
4. Mortality types of known grizzly bear mortalities by year 1997-2002.
5. Mortality types of known grizzly bear mortality by BMA 1997-2002.
6. Gender of known grizzly bear mortalities by year 1997-2002.
7. Ratio of males/females in known grizzly bear mortalities by year 1997-2002.
8. Gender of known grizzly bear mortalities by BMA 1997-2002.
9. Ratios of males/females in known mortalities by BMA 1997-2002.
10. Frequencies of grizzly bear mortalities and removals pooled by BMA and sex.
11. Frequencies of grizzly bear mortalities by pooled BMA categories and sex.
12. Boxplots of change in age and sex of bears obtained from harvest records and translocations.
13. Boxplots of change in age and sex obtained from harvest and relocations.
14. Changes in distribution of ages for male and female bear mortalities for selected WMU's.
15. Trends in harvest success by BMA.
16. Spatial distribution of problem bear captures and relocations.

## Appendix

A-C. Spatial distribution of total grizzly bear mortalities in Alberta.

## Background

In response to the assessment of the Endangered Species Conservation Committee (ESCC), which recommended the immediate closure of the grizzly bear hunting in Alberta, the Minister of Sustainable Resource Development directed the department to conduct a review to allow further consideration of this recommendation.

To conduct this review a small Grizzly Bear Technical Committee was established in July 2002 lead by provincial biologist Gordon Stenhouse. Other committee members included Dr. Mark Boyce (ACA Chair in Fisheries and Wildlife, University of Alberta) and Mr. John Boulanger (Bio-statistician - consultant, Nelson, RC.).

Our committee had the following objectives:

1. Review all the mortality data (sex, age, locations) and translocations (by BMA and WMU) since 1988 and where data exists back to 1973 to explore if any trends or patterns in mortalities are suggested that may be of management concern. The focus here was to determine any biological concerns relating to mortalities and translocations from a population management perspective. A key objective of the analysis was an evaluation of the data as a management tool, and an examination of how the data could be supplemented or augmented with other sources of data.
2. Review the current formula that Alberta Fish and Wildlife Division uses to determine BMA population values. This review focused on assumptions made about grizzly bear population sizes and trends for current allocation of grizzly bear hunting licenses.

## 3. Identify data needs and the resources required to ensure effective management decisions.

## The Data Set

The province of Alberta has kept systematic records of all known grizzly bear mortalities, and translocations since 1971. These records were to include the date, location, sex, type of mortality/translocation, release site locations (where applicable) and a tooth was to be submitted to the department for age determination. The process for registration of legally harvested bears requires the hunter to provide the above information to a local conservation officer who completes a registration form and submits this to department data management staff in Edmonton. In the case of translocated bears or illegal kills the responding conservation officer completes the registrations independently and forwards this information as above.

Previous summaries of provincial grizzly bear mortalities have been compiled for the period 1972-1994 (Gunson 1995), and 1988-1996 (Gunson 1996). Since no comprehensive grizzly bear mortality data set existed in a digital format, we spent a significant amount of effort to assemble a digital data file that incorporated historic data records (from Gunson 1995) as well as more recent annual digital department files into an amalgamated data set. Unfortunately, although each registered grizzly bear mortality or
relocation should have age, sex, and location data, in a large majority of cases some of these data are not available. The largest set of missing values is age of the registered bear. Also, the common practice of recording locations by township and range provides imprecise data when one is interested in the spatial arrangement of grizzly bear mortalities.

The limitations of the available data used in this analysis are important to note and hence some conclusions reached in this document, which address age/sex composition of mortalities, should be viewed with caution as a result of these factors.

For future and ongoing management needs we feel it is important to ensure:

- All grizzly bear mortalities, relocations, and problem bears handled must be registered along with sex, age and precise locations (UTM's). Sufficient resources are needed to ensure that all tooth samples collected are aged in a timely fashion by a recognized laboratory.
- The maintenance of a comprehensive provincial grizzly bear mortality database must be viewed as a high priority within the Department. This database must be continually updated and data entry verified on a regular basis. Effort should be made to deal with database structure and format. Ongoing and regular communication with field staff is necessary to assess the accuracy of annual data.
- Annual summary reports are necessary to monitor the database and evaluate results relative to mortality trends.


## Background on Population Assessment

In 1990 the department published a provincial Grizzly Bear Management Plan, which included the first population estimates of grizzly bears in the province. This plan divided the province into 21 Bear Management Areas (BMA's) (see figure 1) and population estimates were developed for each of these management units. These population estimates were based on habitat-specific bear densities compiled from several intensive grizzly bear research programs. The final densities and hence population estimates were modified further based on a subjective evaluation of the land surface disturbance (human impacts and activities) within a given BMA. This was done in an attempt to integrate a habitat disturbance component into these calculations. Estimated total surface disturbance at that time (1990) averaged $45 \%$. This process resulted in the establishment of the baseline population levels for grizzly bears in Alberta.

Since 1988, the department has completed an annual update to the original 1990 grizzly bear population estimates for each BMA in Alberta. These annual estimates built on the previous years' estimate and incorporated new data on mortalities, translocations, estimated net immigration and net natural growth (Gunson 1996). Historically in Alberta we have defined net natural growth for grizzly bears as allowable human caused mortality at 6\% per year minus known mortality


Figure 1. Bear Management Areas (BMA's) in Alberta

## Comments on Current Formula and Approach used for Harvest Allocation

Table 1 presents a summary of grizzly bear population estimates for each of the provincial BMA's in Alberta from 1988-2002. The numbers presented in this table are calculations derived from the allocation formula used since 1988. Our review of these numbers and the formula itself, suggested a number of changes that would be beneficial in the approach the province uses in setting grizzly bear harvest quotas. First the current model used to estimate population size is unbounded and therefore if no harvest or human-caused sources of mortality are recorded, the population will grow exponentially without bound. This is unrealistic in terms of grizzly bear ecology, and conflicts with the spirit of habitat-based population estimates, which we believe was intended in the 1990 Alberta Grizzly Bear Management Plan.

|  | Table 1: Number of grizzly bears estimated to occur in Alberta Bear Management Areas (BMAs) 1988-2002 (excluding national parks) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| YEAR | BEAR MANAGEMENT AREA (BMA) |  |  |  |  |  |  |  |  |  |  |
|  | 1 | 2A | 2B | 3A | 3B | 4A | 4B | 4C | 6\&7 | REM** | TOTAL |
| 1988 | 82 | 22 | 127 | 79 | 26 | 52 | 34 | 35 | 45 | 73 | 575 |
| 1989 | 84 | 18 | 126 | 85 | 27 | 46 | 30 | 22 | 27 | 71 | 536 |
| 1990 | 90 | 16 | 128 | 88 | 28 | 46 | 35 | 22 | 24 | 70 | 547 |
| 1991 | 90 | 21 | 136 | 86 | 29 | 65 | 43 | 33 | 47 | 88 | 638 |
| 1992 | 92 | 21 | 143 | 88 | 31 | 72 | 46 | 33 | 47 | 96 | 669 |
| 1993 | 93 | 21 | 145 | 97 | 30 | 82 | 45 | 28 | 43 | 102 | 686 |
| 1994 | 96 | 22 | 150 | 101 | 33 | 75 | 45 | 28 | 42 | 108 | 700 |
| 1995 | 102 | 23 | 158 | 108 | 33 | 77 | 44 | 31 | 46 | 113 | 735 |
| 1996 | 109 | 21 | 168 | 112 | 34 | 80 | 46 | 33 | 44 | 118 | 765 |
| 1997 | 119 | 17 | 171 | 108 | 35 | 78 | 47 | 35 | 40 | 126 | 776 |
| 1998 | 124 | 13 | 176 | 120 | 37 | 79 | 5 C | 36 | 31 (57*) | 141 | 807 (833) |
| 1999 | 126 | 10 | 183 | 124 | 44 | 83 | 48 | 37 | 31 (55) | 147 | 833 (857) |
| 2000 | 134 | 7 | 187 | 120 | 47 | 87 | 44 | 35 | 30 (54) | 150 | 841 |
| 2001 | 14C | 12 | 188 | 121 | 48 | 88 | 95 | 36 | 53 | 156 | 940 |
| 2002 | 15C | 8 | 200 | 118 | 52 | 93 | 95 | 36 | 50 | 163 | 968 |

* Numbers in brackets represent revised population estimates taking into account DNA-based population density findings (Mowat and Russel 1998).
**REM = Remainder of Province including BMA 5 (Kananaskis Country), Lougheed Provincial Park, WMU 410 (bow Corridor), BMA 13 and BMA 16.
Note: Table adapted and modified from Kansas 2002.

This concern is relatively simple to rectify. First, we assume that the habitat-based density estimates allow us to calculate a carrying capacity for the landscape. Second, we use this carrying capacity, $K$, as the upper limit on population size without harvest (Table 1 b ). Third, we assume that density dependence kicks in at relatively large population size consistent with Taylor (1994), suggesting a nonlinear density-dependent function. We propose to capture these assumptions with a theta-logistic model with a harvesting rate term, $h$ :

$$
N_{t+1}=N_{t} e^{\left\{0.058\left(1-\left[N_{t} / K\right\}^{3}\right)\right]}-h
$$

where Nt is the grizzly bear population in a BMA in year t .
This model retains the basic elements of the current allocation formula and 1990 Grizzly Bear Management Plan but keeps the population in check relative to its habitats. The harvesting term, $h$, can be calculated using modifications to our current formula with adjustments for net translocations, net immigration, and total known mortality. Thus we use $h=$ hunter kill + nonharvest mortalities - (net immigration + net trans locations).

This model is similar to the catch-limit algorithm used by the International Whaling Commission (Boyce 2000).

Table Ib: Habitat carrying capacity

| BMA | 1988 habitat carrying capacity (K) (bears) |
| :--- | :---: |
| 1 | 82 |
| 2A | 18 |
| 2B | 127 |
| 3A | 79 |
| 3B | 26 |
| 4A | 40 |
| 4B | 28 |
| 4C | 31 |
| 6 and 7 | 29 |

We also believe that change is warranted in the $25 \%$ adjustment figure of unreported mortality currently used in the harvest allocation formula. The inflation of non-harvest mortality is likely to be higher than for hunter kills because poaching and other non-harvest mortality is less likely to be reported. Based on recent research findings, we propose to add a wounding loss of 10-15\%, and the non-harvest mortality should be set at a rate of $100 \%$ (see McLellan et al. 1999). Also, because relocated grizzly bears suffer increased mortality rates, we suggest the net translocations be reduced by $30 \%$. This number should be altered as new research results become available. Overall these changes would improve the utility of this management tool and result in more realistic annual population estimates for grizzly bear populations in Alberta.

We recalculated population estimates using the new formula for 1988 to 2002 (Table 2). One problem that we encountered was that the original estimates of immigration used in calculations prior to 1998 were not available. In addition, the methodology in which immigration was estimated for each BMA (Nagy and Gunson 1990, p. 82» was not detailed enough to allow duplication of the original estimates. Given this, we set immigration to 0 for BMA's with the exception of BMA's 6 and 7 that were adjacent to the Flathead Valley, a documented potential source of dispersal (Hovey and McLellan 1986). For these BMA's net immigration was set to 2 bears per year which was the same value used in calculations with the original formula. Net immigration was set at 0 or 1 bear per year for most BMA's with the original formula, and therefore a large change in estimates should not be expected under the assumption of no immigration into BMA's. It would be useful to have a more rigorous method to estimate grizzly bear dispersals into BMA. For example, genetic methods could be used to estimate population boundaries, . and long-term movement between sub-populations.

The data presented in Figure 1 b represents a recalculation of grizzly bear population estimates using the suggested revised formula for the period 1988-2001. Estimates using the previous formula are also included for comparison purposes. By including a habitat carrying capacity (based on original values used in the 1988 baseline K variable along with the other identified changes) the annual population change does not correspond to
previous calculations and the model does not indicate steady population growth in BMA's 1, 2B, 3A, 3B and 4A, and 6\&7. The old and new formulas indicate similar population size and trends in BMA's 2A and 4C. The population size for the old formula was increased in BMA's 4B, and 6 and 7 in 1990 and 1997 therefore creating further discrepancy between population trajectories.

We emphasize, however, that the $K$ value used in the new formula is one prepared for a 1988 landscape. An assumption of the new formula is that density dependant processes (such as density dependant survival, productivity, and dispersal rates) do not allow the population to exceed K (Figure 1), and therefore the estimate of K will have direct implications of population size in any BMA. One other factor that is not accounted for in this model is emigration and immigration of bears between BMA's. At this time there is no information to allow estimation of these rates. Given lack of information, the most conservative strategy is to assume minimal immigration into BMA's, therefore requiring the productivity of each BMA to offset mortality to maintain population stability.

One other important point to note is that in most BMA's the population remained stable at a level near K. One exception to this was BMA 2A which declined. This suggests that the estimated mortality did not exceed the estimated level of population increase for most BMA's.


Figure lb. Estimates of population size and trend for BMA's using revised and old population models

## Grizzly Bear Allocations

| Habitat |  | A | B C |  | C1 | D | D1 | D2 | D3 | D4 | E |  | G | H | 1 |  | J |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  | 8 | EST. | ALLOW. |  | NET | JAN. 2002 |  |  |
| Carrying |  |  |  |  | Net | EST. |  | Hunt |  | Unreport | Total | total | MAN CAUSED | NATURAL | CAL. |  |  |
|  | acity |  | POP | NET |  | Trans | NET | Hunt | \& wound | Non-hunt | Nórihunt | REPT. | MORT. | MORT. | GROWTH | GRIZ. POP. | Old model | ARV ( |
| Year K |  |  | GBMA | EST. | TRANS. | (-30\%) | MM. | Kill | (+15\%) | Mort. | ( $+100 \%$ ) | MORT. | D2+D4 | @6\%/YR | (G-F) | New model | N | GOAL |
| 1988 | 82 | 1 | 82 | 2 | 1.4 | 0 | 3 | 3.45 | 4 | 8.0 | 7 | 11.5 | 4.9 | -6.53 | 72.0 | 84 | 0.02 |
| 1989 | 82 | 1 | 72.0 | 0 | 0.0 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 4.3 | 3.17 | 85.8 | 90 | 0.02 |
| 1990 | 82 | 1 | 85.8 | 6 | 4.2 | 0 | 6 | 6.9 | 1 | 2.0 | 7 | 8.9 | 5.1 | -3.75 | 74.1 | 90 | 0.02 |
| 1991 | 82 | 1 | 74.1 | 3 | 2.1 | 0 | 2 | 2.3 | 2 | 4.0 | 4 | 6.3 | 4.4 | -1.85 | 82.1 | 92 | 0.02 |
| 1992 | 82 | 1 | 82.1 | 1 | 0.7 | 0 | 4 | 4.6 | 0 | 0.0 | 4 | 4.6 | 4.9 | 0.33 | 78.0 | 93 | 0.02 |
| 1993 | 82 | 1 | 78.0 | 1 | 0.7 | 0 | 3 | 3.45 | 0 | 0.0 | 3 | 3.5 | 4.7 | 1.23 | 81.8 | 96 | 0.02 |
| 1994 | 82 | 1 | 81.8 | 1 | 0.7 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 4.9 | 3.76 | 81.7 | 102 | 0.02 |
| 1995 | 82 | 1 | 81.7 | 1 | 0.7 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 4.9 | 4.90 | 82.9 | 109 | 0.02 |
| 1996 | 82 | 1 | 82.9 | 7 | 4.9 | 0 | 4 | 4.6 | 0 | 0.0 | 4 | 4.6 | 5.0 | 0.38 | 81.6 | 119 | 0.02 |
| 1997 | 82 | 1 | 81.6 | 3 | 2.1 | 0 | 2 | 2.3 | 2 | 4.0 | 4 | 6.3 | 4.9 | -1.41 | 78.1 | 124 | 0.02 |
| 1998 | 82 | 1 | 78.1 | 0.0 | 0.0 | 0 | 4 | 4.6 | 1 | 2.0 | 5 | 6.6 | 4.7 | -1.91 | 77.9 | 126 | 0.02 |
| 1999 | 82 | 1 | 77.9 | 2.0 | 1.4 | 0 | 0 | 0 | 2 | 4.0 | 2 | 4.0 | 4.7 | 0.68 | 82.0 | 134 | 0.02 |
| 2000 | 82 | 1 | 82.0 | 3.0 | 2.1 | 0 | 2 | 2.3 | 2 | 4.0 | 4 | 6.3 | 4.9 | -1.38 | 77.8 | 140 | 0.02 |
| 2001 | 82 | 1 | 77.8 | 4.0 | 2.8 | 0 | 2 | 2.3 | 0 | 0.0 | 2 | 2.3 | 4.7 | 2.37 | 85.2 | 150 | 0.02 |
| 1988 | 18 | 2A | 22 | 0 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.3 | -0.68 | 11.6 | 18 | 0.02 |
| 1989 | 18 | 2A | 11.6 | 0 | 0.0 | 0 | 1 | 1.15 | 1 | 2.0 | 2 | 3.2 | 0.7 | -2.45 | 14.6 | 16 | 0.02 |
| 1990 | 18 | 2A | 14.6 | -2 | 0.0 | 0 | 0 | 0 | 2 | 4.0 | 2 | 4.0 | 0.9 | -3.12 | 15.2 | 21 | 0.02 |
| 1991 | 18 | 2A | 15.2 | -1 | 0.0 | 0 | 0 | 0 | ': | 2.0 | 1 | 2.0 | 0.9 | -1.09 | 17.2 | $\cdots \quad 21$ | 0.02 |
| 1992 | 18 | 2A | 17.2 | -1 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.0 | -0.97 | 16.5 | $\therefore \quad 21$ | 0.02 |
| 1993 | 18 | 2A | 16.5 | 0 | 0.0 | 0 | 0 | 0 | . 0 | 0.0 | 0 | 0.0 | 1.0 | 0.99 | 18.8 | 22 | 0.02 |
| 1994 | 18 | 2A | 18.8 | 0 | 0.0 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 1.1 | -0.02 | 16.1 | 23 | 0.02 |
| 1995 | 18 | 2A | 16.1 | 0 | 0.0 | 0 | 1 | 1.15 | ${ }^{0}$ | 0.0 | 1 | 1.2 | 1.0 | -0.18 | 17.8 | 21 | 0.02 |
| 1996 | 18 | 2A | . 17.8 | -4 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.1 | -0.93 | 16.1 | 17 | 0.02 |
| 1997 | 18 | 2A | 16.1 | -4 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.0 | -1.03 | 17.0 | 13 | 0.02 |
| 1998 | 18 | 2A | 17.0 | 0.0 | 0.0 | 0 | 1 | 1.15 | 2 | 4.0 | 3 | 5.2 | 1.0 | -4.13 | 13.5 | 10 | 0.02 |
| 1999 | 18 | 2A | 13.5 | -2.0 | 0.0 | 0 | 0 | 0 | 1. | 2.0 | 1 | 2.0 | 0.8 | -1.19 | 16.9 | 7 | 0.02 |
| 2000 | 18 | 2A | 16.9 | 0.0 | 0.0 | 0 | 0 | 0 | 5 | 10.0 | 5 | 10.0 | 1.0 | -8.99 | 8.7 | 12 | 0.02 |
| 2001 | 18 | 2A | 8.7 | -3.0 | 0.0 | 0 | 1 | 1.15 | 2 | 4.0 | 3 | 5.2 | 0.5 | -4.63 | 9.4 | 8 | 0.02 |
| 1988 | 127 | 2B | 127 | -1 | 0.0 | 0 | 2 | 2.3 | 0 | 0.0 | 2 | 2.3 | 7.6 | 5.32 | 124.7 | 126 | 0.02 |
| 1989 | 127 | 2B | 124.7 | 5 | 3.5 | 0 | 1 | 1.15 |  | 18.0 | 10 | 19.2 | 7.5 | -11.67 | 113.0 | 128 | 0.02 |
| 1990 | 127 | 2B | 113.0 | 2 | 1.4 | 0 | 2 | 2.3 | 4 | 8.0 | 6 | 10.3 | 6.8 | -3.52 | 125.4 | 136 | 0.02 |
| 1991 | 127 | 2B | 125.4 | 1 | 0.7 | 0 |  | 0 | 2 | 4.0 | 2 | 4.0 | 7.5 | 3.52 | 124.9 | 143 | 0.02 |
| 1992 | 127 | 2B | 124.9 | 1 | 0.7 | 0 | 4 | 4.6 | 3 | 6.0 | 7 | 10.6 | 7.5 | -3.11 | 118.6 | 145 | 0.02 |
| 1993 | 127 | 2B | 118.6 | 0 | 0.0 | 0 | 2 | 2.3 | 1 | 2.0 | 3 | 4.3 | 7.1 | 2.82 | 127.8 | 150 | 0.02 |
| 1994 | 127 | 2B | 127.8 | 5 | 3.5 | 0 | 2 | 2.3 | 3 | 6.0 | 5 | 8.3 | 7.7 | -0.63 | 121.6 | 158 | 0.02 |
| 1995 | 127 | 2B | 121.6 | 3 | 2.1 | 0 | 1 | 1.15 | 1 | 2.0 | 2 | 3.2 | 7.3 | 4.14 | 129.6 | 168 | 0.02 |
| 1996 | 127 | 2B | 129.6 | 3 | 2.1 | 0 | 5 | 5.75 | 2 | 4.0 | 7 | 9.8 | 7.8 | -1.98 | 117.3 | 171 | 0.02 |
| 1997 | 127 | 2B | 117.3 | 2 | 1.4 | 0 | 4 | 4.6 | 2 | 4.0 | 6 | 8.6 | 7.0 | -1.56 | 125.5 | 176 | 0.02 |
| 1998 | 127 | 2B | 125.5 | 0.0 | 0.0 | 0 | 3 | 3.45 | 0 | 0.0 | 3 | 3.5 | 7.5 | 4.08 | 124.6 | 183 | 0.02 |
| 1999 | 127 | 2B | 124.6 | 1.0 | 0.7 | 0 | 4 | 4.6 | 2 | 4.0 | 6 | 8.6 | 7.5 | -1.12 | 120.8 | 187 | 0.02 |
| 2000 | 127 | 2B | 120.8 | -1.0 | 0.0 | 0 | 5 | 5.75 | 2 | 4.0 | 7 | 9.8 | 7.2 | -2.50 | 121.3 | 188 | 0.02 |
| 2001 | 127 | 2B | 121.3 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 7.3 | 7.28 | 130.8 | 200 | 0.02 |
| 1988 | 79 | 3A | 79 | -1 | 0.0 | 0 | 0 | 0 | 2 | 4.0 | 2 | 4.0 | 4.7 | 0.74 | 75.0 | 85 | 0.02 |
| 1989 | 79 | 3A | 75.0 | 1 | 0.7 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 4.5 | 2.50 | 80.3 | 88 | 0.02 |
| 1990 | 79 | 3A | 80.3 | 0 | 0.0 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 4.8 | 3.67 | 76.8 | 86 | 0.02 |
| 1991 | 79 | 3A | 76.8 | -1 | 0.0 | 0 | 2 | 2.3 | 0 | 0.0 | 2 | 2.3 | 4.6 | 2.31 | 78.2 | 88 | 0.02 |
| 1992 | 79 | 3A | 78.2 | 5 | 3.5 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 4.7 | 3.54 | 81.9 | 97 | 0.02 |

## Grizzly Bear Allocations (cont...)

| EAR ALLOCATIONS |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| K | L | M | historic values for m | N | 0 | P | Additional <br> information |
| hUNTER | INTERM LIC | CUMMULATIVE 1972- | CUMMULATIVE 1972-AD | AduSted \#lic. | \# LIC. ALLOCATED | EST. | Licenses alloted Harvest |
| SUJCCESS | ( $\mathbf{x J} \mathrm{J} / \mathrm{K}$ ) | FEMALE MAN | FEMALE MAN | $35 \%$ FEMALE $=$ |  | HARVEST |  |
| (MOST RECENT) | (HARVEST RATE) | CAUSED MORTAL. | CAUSED MORTAL. +TRANS. OUT | RFDUCTION | (DEPENDANT ON REGIONAL REVIEW) |  |  |
| (5 YEARS) | /SUCCESS RATE | +TRANS. OUT |  | OF 0.33 |  |  |  |
|  | 22 | 10 |  | 14 |  |  |  |
|  | 26 | 10 |  | 17 |  |  |  |
| u.v.u | 22 | 14 |  | 15 |  |  |  |
| 6.6\% | 25 | 16 |  | 16 |  |  |  |
| 6.6\% | 24 | 16 |  | ., \% 16 |  |  |  |
| 7.0\% | 23 | 18 |  | 16 |  |  |  |
| 7.4\% | 22 | 18 |  | 14 |  |  |  |
| 5.8\% | 29 | 18 |  | 19 |  |  |  |
| 7.2\% | 23 | 20 |  | 15 |  |  |  |
| 6.3\% | 25 | 22 |  | 16 |  |  |  |
| 6.8\% | 23 | 25 | 33.0 | 15 |  |  | $37 \quad 4$ |
| 5.9\% | 28 | 26 | 34.0 | 18 |  |  | 310 |
| 7.1\% | 22 | 27 | 34.0 | * 14 |  |  | $34 \quad 2$ |
| 6.9\% | 25 | 29 | 36.0 | - 16 |  |  | 17 2 |
| 2.3\% | 10 | 4 |  | 7 |  |  |  |
| 2.3\% | 13 | 5 |  | 8 |  |  |  |
| 2.3\% | 13 | 8 |  | \% 9 |  |  |  |
| 2.3\% | 15 | 8 | $\%$ | - 10 |  |  |  |
| 2.3\% | 15 | 8 | * | 10 |  |  |  |
| 1.8\% | 21 | 8 |  | $\because 14$ |  |  |  |
| 1.3\% | 26 | 8 |  | (i. 17 |  |  |  |
| $\cdots$ - $2.3 \%$ | 15 | 8 | : | 810 |  |  |  |
| 2.3\% | 14 | 9 |  | $\therefore \quad 9$ |  |  |  |
| 2.3\% | 15 | 13 |  | 10 |  |  |  |
| 7.3\% | 4 | 14 | * 42.0 | 2 |  |  | 4 |
| 6.1\% | 6 | 14 | 39.0 | 4 |  |  | 20 |
| 5.0\% | 3 | 17 | 41.0 | 2 |  |  | 20 |
| 7.9\% | 2 | 17 | 38.0 | 2 |  |  | 7 |
| 3.6\% | 69 | 38 |  | 46 |  |  |  |
| 3.6\% | 63 | 45 |  | 41 |  |  |  |
| 3.6\% | 70 | 49 |  | 46 |  |  |  |
| 3.6\% | 69 | 51 |  | 46 |  |  |  |
| - $3.6 \%$ | 66 | 53 |  | 44 |  |  |  |
| 3.8\% | 67 | 53 |  | $\therefore \quad 44$ |  |  |  |
| 4.6\% | 53 | 55 |  | 35 |  |  |  |
| 4.6\% | 57 | 56 |  | 37 |  |  |  |
| 7.2\% | 33 | 58 |  | 21 |  |  |  |
| 8.1\% | 31 | 59 |  | 20 |  |  |  |
| 9.3\% | 27 | 60 | 420 | 18 |  |  | 32 |
| 10.3\% | 23 | 62 | 42.0 | 15 |  |  | 35 |
| 13.9\% | 17 | 65 | 42.0 | $\therefore 11$ |  |  | 24 |
| 11.3\% | 23 | 65 | 42.0 | 15 |  |  | 13 |
| 3.2\% | 47 | 52 |  | 31 |  |  |  |
| 3.2\% | 50 | 54 |  | $\therefore \quad 33$ |  |  |  |
| 3.2\% | 48 | 54 |  | 32 |  |  |  |
| 3.2\% | 49 | 55 |  | 32 |  |  |  |
| 3.2\% | 51 | 57 |  | 34 |  |  |  |

Grizzly Bear Allocations (cont...
GRIZZLY BEAR

| Habitat |  | A | B | C | C1 | D | D1 | D2 | D3 | D4 | E | $\begin{gathered} \mathrm{F} \\ \mathrm{EST} . \end{gathered}$ | G | H | I |  | J |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  |  | ALLOW. | NET | JAN. 2002 |  |  |  |
| Carrying |  |  | POP | NET | Net <br> Trans | EST. |  | Hunt |  | Unreport | Total | TOTAL | MAN CAUSED | NATURAL | CAL. |  | HARV (MC |  |
|  | city | GBMA |  |  |  | NET | Hunt \& | \& wound | Non-hunt | Nonhunt | REPT. | MORT. | MORT. | GROWTH | GRIZ. POP | Old model |  |  |
| Year K |  |  | EST. | TRANS | (-30\%) | IMM | Kill | (+15\%) | Mort. | ( $+100 \%$ ) | MORT. | D2+D4 | @ $6 \% / \mathrm{YR}$ | (G-F) | New model | N | GOAL |  |
| 1993 | 79 | 3A | 81.9 | 0 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 4.9 | 2.92 | 74.6 | 101 | 0.02 |  |
| 1994 | 79 | 3A | 74.6 | 1 | 0.7 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 4.5 | 4.48 | 82.5 | 108 | 0.02 |  |
| 1995 | 79 | 3A | 82.5 | 0 | 0.0 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 5.0 | 3.80 | 74.9 | 112 | 0.02 |  |
| 1996 | 79 | 3A | 74.9 | 0 | 0.0 | 0 | 4 | 4.6 | 2 | 4.0 | 6 | 8.6 | 4.5 | -4.10 | 73.0 | 108 | 0.02 |  |
| 1997 | 79 | 3A | 73.0 | 5 | 3.5 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 4.4 | 4.38 | 86.0 | 120 | 0.02 |  |
| 1998 | 79 | 3A | 86.0 | 2.0 | 1.4 | 0 | 1 | 1.15 | 3 | 6.0 | 4 | 7.2 | 5.2 | -1.99 | 66.8 | 124 | 0.02 |  |
| 1999 | 79 | 3A | 66.8 | 0.0 | 0.0 | 0 | 8 | 9.2 | 1 | 2.0 | 9 | 11.2 | 4.0 | -7.19 | 72.9 | 120 | 0.02 |  |
| 2000 | 79 | 3A | 72.9 | -2.0 | 0.0 | 0 | 2 | 2.3 | 1 | 2.0 | 3 | 4.3 | 4.4 | 0.08 | 78.3 | 121 | 0.02 |  |
| 2001 | 79 | 3A | 78.3 | 0.0 | 0.0 | 0 | 7 | 8.05 | 1 | 2.0 | 8 | 10.1 | 4.7 | -5.35 | 69.5 | 118 | 0.02 |  |
| 1988 | 26 | 3B | 26 | 1 | 0.7 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.6 | -0.44 | 24.7 | 27 | 0.02 |  |
| 1989 | 26 | 3B | 24.7 | -3 | 0.0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 1.5 | 1.48 | 26.8 | 28 | 0.02 |  |
| 1990 | 26 | 3B | 26.8 | -5 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.6 | -0.39 | 23.3 | 29 | 0.02 |  |
| 1991 | 26 | 3B | 23.3 | 0 | 0.0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 1.4 | 1.40 | 27.4 | 31 | 0.02 |  |
| 1992 | 26 | 3B | 27.4 | -3 | 0.0 | 0 | 2 | 2.3 | 0 | 0.0 | 2 | 2.3 | 1.6 | -0.65 | 22.5 | 30 | 0.02 |  |
| 1993 | 26 | 3B | 22.5 | 2 | 1.4 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 1.3 | 0.20 | 27.9 | 33 | 0.02 |  |
| 1994 | 26 | 3B | 27.9 | -6 | 0.0 | 0 | 0 | 0 | 1 | 2.0 | 1 | 2.0 | 1.7 | -0.33 | 22.3 | 33 | 0.02 |  |
| 1995 | 26 | 3B | 22.3 | -1 | 0.0 | 0 | 1 | 1.15 | 0 | 0.0 | 1 | 1.2 | 1.3 | 0.19 | 26.5 | 34 | 0.02 |  |
| 1996 | 26 | 3B | 26.5 | -1 | 0.0 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 1.6 | 1.59 | 25.6 | 35 | 0.02 |  |
| 1997 | 26 | 3B | 25.6 | 5 | 3.5 | 0 | 2 | 2.3 | 2 | 4.0 | 4 | 6.3 | 1.5 | -4.76 | 23.5 | 37 | 0.02 |  |
| 1998 | 26 | 3B | 23.5 | 5.0 | 3.5 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 1.4 | 1.41 | 30.9 | 44 | 0.02 |  |
| 1999 | - 26 | 3B | 30.9 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | - 0.0 | 0 | 0.0 | 1.9 | 1.85 | 20.8 | 47 | 0.02 |  |
| 2000 | 26 | 3B | 20.8 | 1.0 | 0.7 | 0 | 1 | 1.15 | 1 | 2.0 | 2 | 3.2 | 1.2 | -1.90 | 25.2 | - 48 | 0.02 | - |
| 2001 | 26 | 3B | 25.2 | 1.0 | 0.7 | 0 | 0 | 0 | 0 | - 0.0 | 0 | 0.0 | 1.5 | 1.51 | 27.2 | 52 | 0.02 |  |
| 1988 | 40 | 4A | 52 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | ) 0.0 | 0 | 0.0 | 3.1 | 3.12 | 25.9 | 46 | 0.02 |  |
| 1989 | 40 | 4A | 25.9 | -3 | 0.0 | 0 | 0 | 0 | 0 | ) 0.0 | 0 | 0.0 | 1.6 | 1.55 | 39.6 | 46 | 0.02 |  |
| 1990 | 40 | 4A | 39.6 | 1 | 0.7 | 0 | 2 | 2.3 | 1 | 2.0 | 3 | 4.3 | 2.4 | -1.92 | 36.7 | 65 | 0.02 |  |
| 1991 | 40 | 4A | 36.7 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | 0 0.0 | 0 | 0.0 | 2.2 | 2.20 | 41.9 | 72 | 0.02 |  |
| 1992 | 40 | 4A | 41.9 | 6 | 4.2 | 0 | 2 | 2.3 | 0 | ) 0.0 | 2 | 2.3 | 2.5 | 0.21 | 40.3 | 82 | 0.02 |  |
| 1993 | 40 | 4A | 40.3 | 0 | 0.0 | 0 | 6 | 6.9 | 5 | 510.0 | 11 | 16.9 | 2.4 | -14.48 | 22.9 | 75 | 0.02 |  |
| 1994 | 40 | 4A | 22.9 | 0 | 0.0 | 0 | 1 | 1.15 | 2 | 24.0 | 3 | 5.2 | 1.4 | -3.78 | 31.6 | 77 | 0.02 |  |
| 1995 | 40 | 4A | 31.6 | 1 | 0.7 | 0 | 2 | 2.3 | 0 | 00.0 | 2 | 2.3 | 1.9 | -0.40 | 40.9 | 80 | 0.02 |  |
| 1996 | 40 | 4A | 40.9 | -2 | 0.0 | 0 | 2 | 2.3 | 1 | 12.0 | 3 | 4.3 | 2.5 | -1.85 | 35.0 | 78 | 0.02 |  |
| 1997 | 40 | 4A | 35.0 | -2 | 0.0 | 0 | 0 | 0 | 1 | 12.0 | 1 | 2.0 | 2.1 | 0.10 | 40.4 | 79 | 0.02 |  |
| 1998 | 40 | 4A | 40.0 | 0.0 | 0.0 | 0 | 1 | 1.15 | 0 | $0 \quad 0.0$ | 1 | 1.2 | 2.4 | 1.25 | 38.9 | 83 | 0.02 |  |
| 1999 | 40 | 4A | 38.9 | 0.0 | 0.0 | 0 | 1 | 1.15 | 0 | $0 \quad 0.0$ | 1 | 1.2 | 2.3 | 1.18 | 39.6 | 87 | 0.02 |  |
| 2000 | 40 | 4A | 39.6 | 0.0 | 0.0 | 0 | 3 | 3.45 | 0 | $0 \quad 0.0$ | 3 | 3.5 | 2.4 | -1.07 | 36.8 | 88 | 0.02 |  |
| 2001 | 40 | 4A | 36.8 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | 00.0 | 0 | 0.0 | 2.2 | 2.21 | 41.9 | 93 | 0.02 |  |
| 1988 | 28 | 4B | 34 | 1 | 0.7 | 0 | 0 | 0 | 0 | 0.0 | 0 | 0.0 | 2.0 | 2.04 | 22.1 | 30 | 0.02 |  |
| 1989 | 28 | 4B | 22.1 | 4 | 2.8 | 0 | 2 | 2.3 | 0 | $0 \quad 0.0$ | 2 | 2.3 | 1.3 | -0.97 | 30.2 | 35 | 0.02 |  |
| 1990 | 28 | 4B | 30.2 | 0 | 0.0 | 0 | 2 | 2.3 | 0 | $0 \quad 0.0$ | 2 | 2.3 | 1.8 | -0.49 | 23.7 | 43 | 0.02 |  |
| 1991 | 28 | 4B | 23.7 | 3 | 2.1 | 0 | 3 | 3.45 | 0 | 0.0 | 3 | 3.5 | 1.4 | -2.03 | 28.5 | 46 | 0.02 |  |
| 1992 | 28 | 4B | 28.5 | 0 | 0.0 | 0 | 0 | 0 | 1 | 12.0 | 1 | 2.0 | 1.7 | -0.29 | 25.6 | 45 | 0.02 |  |
| 1993 | 28 | 4B | 25.6 | 0 | 0.0 | 0 | 2 | 2.3 | 1 | 1.2 .0 | 3 | 4.3 | 1.5 | -2.76 | 25.1 | 45 | 0.02 |  |
| 1994 | 28 | 4B | 25.1 | 0.0 | 0.0 | 0 | 0 | 0 | 0 | 00.0 | 0 | 0.0 | 1.5 | 1.50 | 29.6 | 44 | 0.02 |  |
| 1995 | 28 | 4B | 29.6 | 1 | 0.7 | 0 | 2 | 2.3 | 2 | 24.0 | 4 | 6.3 | 1.8 | -4.53 | 21.1 | 46 | 0.02 |  |
| 1996 | 28 | 4B | 21.1 | 1 | 0.7 | 0 | 2 | 2.3 | 1 | 12.0 | 3 | 4.3 | 1.3 | -3.04 | 25.8 |  | . 0.02 |  |
| 1997 | 28 | 4B | 25.8 | 2 | 1.4 | 0 | 0 | 0 | 0 | 00.0 | 0 | 0.0 | 1.6 | 1.55 | 30.7 | 50 | 0.02 |  |

## Grizzly Bear Allocations (cont...)

ALLOCATIONS

| K | L | M | historic values for $m$ | N | 0 | P |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| HUNTER | INTERM LIC | CUMMULATIVE 1972- | CUMMULATIVE 1972- | ADJUSTED \#LIC. | \# LIC. ALLOCATED | EST. | License | Irrest |
| SUCCESS | ( $\mathrm{IxJ} / \mathrm{K}$ ) | FEMALE MAN | FEMALE MAN | >35\% FEMALE $=$ |  | HARVEST | , |  |
| ISTRECENT) | (HARVEST RATE) | CAUSED MORTAL. | CAUSED MORTAL. | REDUCTION | (DEPENDANT ON |  |  |  |
| 5 YEARS) | ISUCCESS RATE | +TRANS. OUT | +TRANS. OUT | OF 0.33 | REGIONAL REVIEW) |  |  |  |
| 3.2\% | 46 | 57 |  | 31 |  |  |  |  |
| 3.2\% | 51 | 57 |  | 34 |  |  |  |  |
| 3.2\% | 47 | 57 | - | 31 |  |  |  |  |
| 2.9\% | 50 | 60 |  | 33 |  |  |  |  |
| 2.1\% | 83 | 60 |  | c. 55 |  |  |  |  |
| 2.5\% | 53 | 61 | 36.0 | - 35 |  |  | 45 | 1 |
| 5.4\% | 27 | 63 | 35.0 | 18 |  |  | 55 | 8 |
| 6.0\% | 26 | 66 | 36.0 | 17 | . |  | 40 | 2 |
| 8.9\% | 16 | 67 | 35.0 | 10 |  |  | 31 | 7 |
| 3.3\% | 15 | 2 |  | 10 |  |  |  |  |
| 3.3\% | 16 | 3 |  | 11 |  |  |  |  |
| 3.3\% | 14 | 8 |  | 9 |  |  |  |  |
| 3.3\% | 16 | 8 |  | 11 |  |  |  |  |
| 3.3\% | 13 | 9 |  | 9 |  |  |  |  |
| 4.2\% | 13 | 11 |  | - 9 |  |  |  |  |
| 4.2\% | 11 | 15 |  | 7 |  |  |  |  |
| 6.2\% | 9 | 16 |  | - 6 |  |  |  |  |
| 6.2\% | 8 | 16 |  | - 5 |  |  |  |  |
| 4.9\% | 10 | 17 |  | i 6 |  |  |  |  |
| 4.1\% | 15 | 17. | 16.0 | ( 10 |  |  | 11 | 0 |
| 4.1\% | 10 | 17 | 16.0 | 17 |  |  | 17 | 0 |
| - $3.1 \%$ | 16 | 18 | 19.0 | - 11 |  |  | 21. |  |
| 3.1\% | 18 | 18 | 19.0 | 12 | . . . |  | 25 | 0 |
| 3.7\% | 14 | 53. |  | - 9 |  |  | . |  |
| 3.7\% | 22 | 54 |  | 14 |  |  |  |  |
| 3.7\% | 20 | 54 |  | - 13 |  |  |  |  |
| 3.7\% | 23 | 54 |  | - 15 |  |  |  |  |
| 3.7\% | 22 | 56 |  | - 15 |  |  |  |  |
| 7.5\% | 6 | 59 |  | 4 |  |  |  |  |
| 9.8\% | 6 | 62 |  | 4 | ? |  |  |  |
| 13.1\% | 6 | 62 |  | - 4 | , 2* |  |  |  |
| 18.1\% | 4 | 64 |  | 3 | - |  |  |  |
| 16.1\% | 5 | 65 |  | 3 | \%-s) |  |  |  |
| 15.6\% | 5 | 66 | 39.0 | 3 |  |  | 6 | 1 |
| 16.2\% | 5 | 67 | 40.0 | 3 |  |  | 7 | 1 |
| 19.8\% | 4 | 67 | 39.0 | 2 |  |  | 7 | 3 |
| 14.8\% | 6 | 67 | 39.0 | 4 |  |  | 5 | 0 |
| 7.3\% | 6 | 10 |  | 4 |  |  |  |  |
| 7.3\% | 8 | 11 |  | 5 |  |  |  |  |
| 7.3\% | 7 | 11 |  | 4 |  |  |  | . |
| 7.3\% | 8 | 12 |  | 5 |  |  |  |  |
| 7.3\% | 7 | 13 |  | 5 |  |  |  |  |
| 10.4\% | 5 | - 14 |  | 3 |  |  |  |  |
| 8.1\% | 7 | -14 |  | 5 |  |  |  |  |
| 8.6\% | 5 | 16 |  | 3 |  |  |  |  |
| 8.2\% | 6 | 16 |  | 4 |  |  |  |  |
| 8.2\% | 8 | 16 |  | 5 |  |  |  |  |

## Grizzly Bear Allocations (cont...)

GRIZZLY Bl


## Grizzly Bear Allocations (cont...)

ミAR ALLOCATIONS


Another important weakness in the harvest allocation model that we feel must be addressed is the \% disturbance on the landscape, which directly affects the estimated grizzly bear population size. In the original calculations of landscape disturbance presented in the 1990 Grizzly Bear Management Plan, the procedures and methods used to make these calculations are unclear. The amount and level of landscape change in grizzly bear habitat in Alberta has changed substantially since the late 1980's and likely will continue to do so. As of this date, the amount of landscape development that has occurred has not been reviewed and included in the department's annual grizzly bear population assessment since 1990 (or in the example calculations presented in Table 2). The ESCC status report and recommendations to the Minister have clearly identified that perceived loss of habitat and pressures on existing grizzly bear habitats are key factors in the conclusions presented, thus it appears that a review and reassessment of this issue is necessary. This is not a simple or quick task and it has not been undertaken as part of this report. Should this be undertaken it is important to formalize this process and ensure that the analysis has statistical rigor. We suggest that a BMA with the most GIS land-use information and grizzly bear data be selected as a pilot area in which to conduct this assessment.

Our review of the data set used to determine the annual population status of grizzly bears in Alberta also pointed out one other important issue relating to model "recalibration " or "resetting". It appears that in the past population estimates for specific BMA's were adjusted by either a "Delphi" (expert opinion) approach of through extrapolation of research results. Although the view that "some data is better that none" and that "those in bear habitat really have a true sense of what bear populations are really doing", these are questionable practices and are not scientifically defensible. We believe that standards must be established for modifications and or adjustments of population numbers. In the case of extrapolation from DNA census results we suggest that, after a careful peer review of the DNA finding the numbers used are restricted to using the lower end of the $50 \%$ confidence interval. This approach ensures a conservative approach to management decisions concerning this species.

The consequences of error in population management for grizzly bears are high as this species reproduces slowly and reduced populations require many years to recover. Miller (1989) using simulation modeling showed that when reproductive rates were generous, natural mortality rates were low, and harvests were $75 \%$ of maximum sustained rates, grizzly bear populations that had been reduced by half would require $>40$ years to recover. Further, even when hunting rates were reduced to $50 \%$ and $0 \%$ during the recovery period, it would take 19 and 10 years respectively for populations to recover from a $50 \%$ reduction in abundance.

## Summary of Known Grizzly Bear Mortalities in Alberta from 1997 to 2002.

Known grizzly bear mortalities in Alberta from 1997 to 2002 were analysed for this summary using comparisons among years and BMA's (Table 3). These numbers represent only known mortalities and likely should be adjusted upwards as identified previously under suggested model adjustments. Averages of the 6 years analysed were compared with the previous 8 years (19881996).

Total known grizzly bear mortalities in Alberta from all causes averaged 29.2 from 1997 to 2002 whereas the average for the period 1988-1996 was 21 bears. Total known mortalities ranged from 20 in 1997 to 38 in 1999 (Figure 2).


Figure 2. Known grizzly bear mortalities by year, 1997-2002.
When these mortalities are broken down by bear management units, GBMAs 1, 2B, 3A, 4B and 7 had relatively high mortality rates ranging from 17 to 26 . GBMAs 2 A and 4 C had mortality rates averaged over 6-years of 14 . GBMAs 3B, 4A, 5, and 6 had relatively low mortality rates of between 7 and 9 while GBMAs 12, 13, and 16 had very low mortality rates between 1 and 3 (Figure 3). Those GBMAs with high or moderate levels of mortality are in areas with large urban communities (i.e. Hinton and Grande Prairie) or extensive agricultural development in association with forested land or protected areas such as Waterton Lakes and Jasper National Parks. Those GBMAs with very low levels of mortality are in areas that are outside of or on the edge of currently identified grizzly bear range where bear densities are low. GBMA 4A has low levels of mortality, possibly due to the lack of motorized access into this area.

Table 3.Total known mortality of grizzly bears by Alberta Bear Management Areas (BMAs) 1972-2002

| YEAR | BEAR MANAGEMENT AREA (BMA) |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2A | 2B | 3A | 3B | 4A | 4B | 4C | 6\&7 | REM*** | TOTAL |
| 1972 | 1 | 0 | 1 | 5 | 1 | 6 | 2 | 2 | 0 | 0 | 18 |
| 1973 | 1 | 0 | 0 | 10 | 0 | 1 | 1 | 2 | 0 | 2 | 17 |
| 1974 | 2 | 0 | 2 | 8 | 1 | 3 | 1 | 0 | 7 | 2 | 26 |
| 1975 | 1 | 1 | 3 | 7 | 0 | 6 | 2 | 0 | 0 | 1 | 21 |
| 1976 | 1 | 2 | 5 | 10 | 1 | 8 | 2 | 6 | 2 | 1 | 38 |
| 1977 | 3 | 1 | 6 | 6 | 1 | 8 | 3 | 4 | 0 | 0 | 32 |
| 1978 | 0 | 2 | 6 | 6 | 1 | 3 | 4 | 4 | 0 | 1 | 27 |
| 1979 | 1 | 1 | 4 | 8 | 1 | 7 | 1 | 5 | 4 | 1 | 33 |
| 1980 | 1 | 1 | 5 | 4 | 1 | 11 | 3 | 11 | 0 | 1 | 38 |
| 1981 | 5 | 1 | 6 | 8 | 2 | 7 | 2 | 12 | 1 | 2 | 46 |
| 1982 | 3 | 2 | 11 | 12 | 1 | 6 | 2 | 5 | 5 | 1 | 48 |
| 1983 | 6 | 2 | 7 | 9 | 2 | 14 | 0 | 12 | 8 | 0 | 60 |
| 1984 | 3 | 1 | 10 | 7 | 1 | 15 | 6 | 6 | 9 | 1 | 59 |
| 1985 | 1 | 1 | 9 | 9 | 0 | 18 | 5 | 6 | 5 | 1 | 55 |
| 1986 | 3 | 1 | 10 | 6 | 0 | 12 | 6 | 6 | 7 | 0 | 51 |
| 1987 | 2 | 4 | 11 | 2 | 2 | 10 | 6 | 12 | 14 | 5 | 68 |
| 1988 | 6 | 1 | 2 | 2 | 0 | 0 | 0 | 2 | 4 | 0 | 17 |
| 1989 | 1 | 2 | 9 | 1 | 0 | 0 | 2 | 1 | 2 | 2 | 20 |
| 1990 | 7 | 2 | 6 | 1 | 1 | 3 | 2 | 6 | 4 | 0 | 32 |
| 1991 | 4 | 1 | 2 | 2 | 0 | 0 | 3 | 2 | 2 | 1 | 17 |
| 1992 | 4 | 1 | 7 | 1 | 2 | 2 | 1 | 6 | 4 | 2 | 30 |
| 1993 | 2 | 1 | 2 | 2 | 1 | 11 | 3 | 2 | 5 | 1 | 30 |
| 1994 | 1 | 1 | 5 | 0 | 1 | 3 | 0 | 0 | 2 | 2 | 15 |
| 1995 | 0 | 1 | 2 | 2 | 1 | 2 | 2 | 1 | 2 | 1 | 14 |
| 1996 | 3 | 3 | 7 | 6 | 0 | 3 | 3 | 2 | 1 | 0 | 28 |
| 1997 | 4 | 1 | 6 | 0 | 4 | 1 | 0 | 1 | 3 | 0 | 20 |
| 1998 | 4 | 3 | 3 | 4 | 0 | 1 | 2 | 1 | 2 | 3 | 23 |
| 1999 | 1 | 1 | 6 | 9 | 0 | 1 | 5 | 3 | 5 | 5 | 36 |
| 2000 | 4 | 5 | 7 | 3 | 2 | 3 | 4 | 1 | 3 | 3 | 35 |
| 2001 | 2 | 2 | 0 | 8 | 0 | 0 | 4 | 2 | 5 | 0 | 23 |
| 2002 | 2 | 2 | 2 | 2 | 1 | 3 | 8 | 6 | 5 | 2 | 33 |



Figure 3. Known grizzly bear mortalities by GBMA, 1997-2002. Averaged over 6 years.
Human-caused mortality was broken down into 3 groups, legal harvest, illegal kills and other mortality (Figure 4). Other mortality included self-defence, research losses, problem wildlife, treaty Indian, and found dead. Legal harvest comprised the largest portion of mortality causes in each of the 6 years analysed, ranging from 10 in 1997 to 19 in 2000 and averaging 15.2. Those mortality causes grouped under other mortalities comprised the second most-common cause of mortality, ranging from 5 in 2001 to 14 in 1999 and averaging 8.3. Illegal kills ranged from 3 in 1998 to 9 in 2002, averaging 5.7. Again, there did not appear to be a trend, either increasing or decreasing, during the 6 years analysed.


Figure 4. Mortality types of known grizzly bear mortalities by year, 1997-2002.

Legal harvest comprised the largest portion of known mortalities in most GBMAs (1, 2B, 3A, 3B, 4A, 4B, 4C, 6 and 7) (Figure 4). GBMAs 2B, 3A, and 4A had legal harvest greater than twice the combined mortality from all other causes. All of these GBMAs are in west central Alberta and include the Willmore Wilderness. Outside of the Willmore, these areas comprise industrial forests on crown lands with a high density of access. These areas, except the Willmore, also have the largest number of mortalities of all GBMAs. In GBMAs 2B and 3A, good access on crown lands probably attracts larger numbers of legal hunters and allows a higher success rate (1996-2000 average $=17.0 \%$ and $11.8 \%$ for GBMA 2 B and 3 A respectively compared with $6 \%$, $5 \%, 2.6 \%, 0 \%, 7.8 \%$, and $0 \%$ for GBMAs I, 2A, 3B, 4A, 4B, and 4C respectively. GBMA 6 and 7 combined also had high hunter success at 10.6\%)). Legal harvests in these 2 units are significantly higher than the average (9.1) for all units (not including GBMA 12, 13, and 16, which are outside of or on the edge of normal grizzly bear range). Whereas illegal kills and other mortalities for these 2 units were close to average.


Figure 5. Mortality types of known grizzly bear mortalities by GBMA, 1997 - 2002. Averaged over 6 years.

Male grizzly bears comprised the largest portion of all known mortalities in each of the years analysed, ranging from 12 in 1997 to 23 in 2000 and averaging 17.7 (Figure 6). The female portion of the known mortalities ranged from 6 in 1997 and 2001 to 16 in 1999 averaging 10.7. In 2002, the number of dead males exceeded dead females by only 2 animals. This high ratio of females in the total human-caused mortality in 2002 is a serious management concern.


Figure 6. Gender of known grizzly bear mortalities by year, 1997-2002.
The ratios of male mortalities to female mortalities (males/females) ranged from 53/47 in 2002 to 76/24 in 2001 (Figure 7). The average over the 6 years was 63/37.The data did not appear to represent any trend in the ratios. The Management Plan for Grizzly Bears in Alberta has a goal of that females should not represent more than $35 \%$ of the annual and cumulative known mortality in each GBMA. The proportion of females among known mortalities exceeded the management goal on a provincial level in 1998, 1999, and 2002 although the average ratio for the 6 years was within the goal.


Figure 7. Ratio of males and females in known grizzly bear mortalities by year, 19972002.

Males comprised the largest portion of known mortalities in all GBMAs (averaged over 6 years) except 1 and 5 (Figure 8). Number of male deaths exceeded females by only 1 and 2 animals in GBMAs 4A and 4B respectively. GBMAs 12, 13 and 16 had no female mortality recorded. The mortality in these units may represent dispersing young males.


Figure 8. Gender of known grizzly bear mortalities by GBMA, 1997-2002. Averaged over 6 years.

Averaged over 6 years, GBMAs 1, 2A, 4A, 4B, and 5 had mortality ratios exceeding the goal of $65 \%$ males to $35 \%$ females or less (Figure 9). The ratio in GBMA 2A was only slightly higher than the goal. Ratios in GBMAs 1, 4A, 4B, and 5 were significantly greater than the goal. GBMAs 1 and 5 had ratios in which the proportion of females exceeded males (39/61 and 4,3/57 respectively). Ratios for all GBMAs (except 12, 13, and 16) ranged from 39/61 to 75/25.


Figure 9. Ratios of males and females in known grizzly bear mortalities by GBMA, 1997 2002. Averaged over 6 years.

Because the loss of adult female grizzly bears has the greatest impact on population, attempts should be made to reduce the number of female grizzly bears in the total human caused mortality (Eberhardt et al. 1994). One possible method to accomplish this is to shorten the length of the hunting season because female bears are more likely to be killed during the latter part of the hunting season (Anon 1990). In addition more effort may be required by SRD staff to keep female grizzly bears within their ecosystem as part of problem bear management actions.

## Additional Analysis of Age/Sex Data

Since grizzly bear mortality rates are generally considered to be low in number on an annual basis, the ability to make inferences about the population status from mortality data only is difficult (Table 3). Further, we know little about the time required for harvest (or total mortality) to impact the population's sex and age structure. In reality, the detection of bear population trend from the sex and/or age structure of harvested bears is a difficult and tenuous task. Changes in population structure indicate a change in a host of variables including food supply, reproduction, mortality or all of these.

Other data are needed to identify the source of population change. Unfortunately, obtaining actual estimates of bear numbers, densities, and trends is difficult, expensive and requires a longterm ongoing program to obtain meaningful management results. In the absence of other data sets it is important to undertake bear management strategies in a conservative manner.

In general terms, the combined use of age and sex information of harvest (and possibly all mortalities) is a stronger approach than simply using age structure alone. However, with sparse data sets encountered with most grizzly bear demographic data relating to
mortalities it is often necessary to investigate the sex ratio of harvest alone. . This approach allows for a significant increase in sample size when one conducts an analysis by BMA. It is important to emphasize that when birth and death rates are constant, the sex and age composition of the population will stabilize regardless of the population trend. Therefore a population could be undergoing a serious decline ( or increase) yet this may not be apparent in a review of the sex/age data of the known mortalities.

With these caveats and for the purpose of this report we have tried to identify possible trends apparent in the available data to assist in pressing management decisions.

We undertook an analysis of change in sex ratio with age over time with the Alberta grizzly bear mortality data set. One major problem we faced was the small number of grizzly bear mortalities in each BMA and that only a small proportion of dead bears were aged.

In an effort to overcome the sample size problem we elected to combine some adjacent BMA 's into single units for the following analysis (see table 4)

Table 4. Grizzly Bear Management Areas Combined for Sex/Age Analysis Purposes

| $1+16$ | $=1$ |
| :---: | :---: |
| $2 \mathrm{~A}+2 \mathrm{~B}$ | $=2$ |
| $3 \mathrm{~A}+3 \mathrm{~B}+13$ | $=3$ |
| $4 \mathrm{~A}+4 \mathrm{~B}+4 \mathrm{C}$ | $=4$ |
| $5+6+7$ | $=5$ |

We looked at the observed change in sex ratio with age over time (Paloheimo and Fraser 1981; Fraser et al. 1982) to estimate harvest rate and other parameters. The use of this method to estimate harvest rate is controversial (Harris and Metzgar 1987) but we felt it has some utility as a trend index.

One obvious problem with this technique when applied to the current data set is sparse sample sizes in terms of number of bear mortalities and removals in each study area. In addition to pooling BMA's as noted above we also combined years to increase the sample sizes (into 4 year blocks 1986-89, 1990-3, 1994-97, 1998-2002). Sample sizes for the pooled data are shown in Figure 10 and 11 below.


Figure 10: Frequencies of grizzly bear mortalities and removals by pooled GBMA categories and sex.


Figure 11: Frequencies of grizzly bear mortalities by pooled GBMA categories and sex.
These summary statistics suggest the greatest number of mortalities and removals occurred in grizzly bear management area 4. In addition, there was no marked change in sex ratio of harvest or removals over time.

## Trends in sex and age ratios over time

We also looked at the change in median age of each sex of bear over time. The rationale behind this method is that a shift in mortalities towards younger age classes should occur if a population is being affected by harvest or other factors (i.e. habitat loss, etc). Basically, the premise would be that all the older bears are being removed (due to higher mortality rates) shifting the distribution of age to younger bears (under the assumption of a constant reproductive rate). In some cases this analysis used individual BMA's while in other cases pooling was necessary due to sparse data (Table 5). Since little difference was found between analysis using mortality data separately and the pooled with removal data as well only the pooled results are presented here.

Table 5: Frequencies of samples aged and not aged for removal and mortality data by pooled years of analysis and GBMA's.

| GBMA(s) | Years | Aged ${ }^{1}$ | Not aged | GBMA(s) | Years | Aged ${ }^{1}$ | Not aged |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1\&16 | 1986-89 | 10 (2) | 3 | 4A | 1986-89 | 23 (8) | 2 |
| 1\&16 | 1990-93 | 17 (7) | 1 | 4A | 1990-93 | 19 (5) | 0 |
| 1\&16 | 1994-97 | 8 (4) | 1 | 4A | 1994-98 | 10 (5) | 2 |
| 1\&16 | 1998-02 | 14 (8) | 3 | 4A | 1999-02 | 6 (2) | 3 |
| 2A | 1986-89 | 6 (3) | 1 | 4B | 1986-89 | 15 (3) | 0 |
| 2A | 1990-93 | 5 (1) | 3 | 4B | 1990-93 | 8 (2) | 1 |
| 2A | 1994-97 | 11 (5) | 0 | 4B | 1994-98 | 5 (1) | 2 |
| 2A | 1998-02 | 11 (2) | 9 | 4B | 1999-02 | 17 (6) | 16 |
| 2B | 1986-89 | 29 (15) | 3 | 4 C | 1986-89 | 22 (11) | 4 |
| 2B | 1990-93 | 16 (8) | 1 | 4C | 1990-93 | 17 (4) | 3 |
| 2B | 1994-97 | 18 (6) | 2 | 4C | 1994-98 | 5 (0) | 1 |
| 2B | 1998-02 | 15 (4) | 5 | 4C | 1999-02 | 7 (2) | 5 |
| 3A | 1986-89 | 14 (9) | 0 | 5\&6 | 1986-89 | 18 (6) | 3 |
| 3A | 1990-93 | 10 (3) | 0 | 5\&6 | 1990-93 | 22 (10) | 0 |
| 3A | 1994-97 | 7 (2) | 3 | 5\&6 | 1994-98 | 9 (3) | 4 |
| 3A | 1998-02 | 22 (7) | 6 | 5\&6 | 1999-02 | 6 (2) | 11 |
| 3B\&13 | 1986-89 | 9 (3) | 2 | 7 | 1986-89 | 19 (4) | 2 |
| 3B\&13 | 1990-93 | 14 (8) | 0 | 7 | 1990-93 | 10 (1) | 2 |
| 3B\&13 | 1994-97 | 13 (5) | 3 | 7 | 1994-98 | 28 (11) | 3 |
| 3B\&13 | 1998-02 | 3 (0) | 3 | 7 | 1999-02 | 40 (14) | 6 |
|  |  |  |  | 12 | 1999-02 | 0 (0) | 1 |

${ }^{1}$ Sample sizes of females aged in parenthesis

Figure 12 displays the results for GBMA's in which a trend in female age over time was suggested. In review, for boxplots, the points connected by the lines are the median. The box are the 25th and 75th percentiles. The bars indicate the range of the data to a maximum of the 1.5 inter-quartile ranges. If an observation is beyond this (a suggested outlier) then a plot symbol is drawn. The inter-quartile range is the distance between the 25th and 75th percentile as indicated by the width of the box. Where no box is shown this is likely a result of spare data (i.e. 1 data point)

The y-axis is on the same scale to allow comparison in ages of bears harvested in all units. One interesting results is that very few older bears were harvested in some GBMA's (2A and 4C). The results for pooled GBMA's 1 and 16 mainly pertain to GBMA 1 since only 2 aged mortalities were present in GBMA 16.


Figure 12: Boxplots of change in age and sex of bears obtained from harvest and removals. A decrease in female age over time is indicated by these figures. Axes are staggered for sexes to prevent boxes from overlapping.


Figure 13: Boxplots of change in age and sex of bears obtained from harvest and removals. No trend in age is evident in these figures.

As previously mentioned we must emphasize that this analysis has limited power to detect changes in population status. One of the main assumptions of this analysis is that the sample of bears harvested represents a random sample of the population. This assumption could easily be violated if younger bears are more vulnerable to harvest. Recent research findings from Stenhouse and Munro (2001) clearly indicate that female bears spend more time in closer proximity to roads thus increasing their risk of mortality. Results from this analysis suggest that BMA's (2B, 4B and 4C) which show a trend in declining age structure of female bears over time may warrant further investigation using more precise techniques. In a conservative management approach it would be prudent to take steps to reduce overall man caused mortality in these BMA's until better population inventory data is available.

Analysis of trends in age and sex in wildlife management units that experienced high historic mortality and problem wildlife removal rates

The previous analyses considered changes in ages of bears for grizzly bear management areas. Further scrutiny has revealed that mortalities within grizzly bear management units are seldom uniform across the area (see figures A, B, C in appendix). In particular; the areas listed in Table 6 have experienced high historical (prior to 1986) and higher mortality numbers in recent years (in some cases).

Table 6: Mortalities, management removals, and frequencies of aged bears for selected wildlife management units

| GBMA | WMU | Years | Aged | Not aged | Total <br> mortalities |
| :--- | :--- | :---: | :---: | :---: | :---: |
| 3A | 350 | $1972-1986$ | 35 | 8 | 43 |
|  | 350 | $1986-1989$ | 6 | 0 | 6 |
|  | 350 | $1990-1993$ | 6 | 0 | 6 |
|  | 350 | $1994-1997$ | 1 | 0 | 1 |
|  | 350 | $1998-2002$ | 1 | 1 | 2 |
| 2B | $354 \& 356$ | $1972-1986$ | 65 | 10 | 75 |
|  | 3548356 | $1986-1989$ | 27 | 3 | 30 |
|  | $354 \& 356$ | $1990-1993$ | 14 | 1 | 15 |
|  | $354 \& 356$ | $1994-1997$ | 15 | 4 | 19 |
|  | $354 \& 356$ | $1998-2002$ | 9 | 2 | 11 |
| 6 and 7 | $402,400, \& 300$ | $1972-1986$ | 29 | 5 | 34 |
|  | $402,400, \& 300$ | $1986-1989$ | 30 | 2 | 32 |
|  | $402,400, \& 300$ | $1990-1993$ | 22 | 1 | 23 |
|  | $402,400, \& 300$ | $1994-1997$ | 16 | 5 | 21 |
|  | $402,400, \& 300$ | $1998-2002$ | 30 | 4 | 34 |
|  | 524 | $1972-1986$ | 20 | 1 | 21 |
|  | 524 | $1986-1989$ | 8 | 3 | 11 |
|  | 524 | $1990-1993$ | 10 | 0 | 10 |
|  | 524 | $1994-1997$ | 6 | 2 | 8 |
|  | 524 | $1998-2002$ | 12 | 2 | 14 |

Some wildlife management units, such as 440 in GBMA 4A experienced high mortality and removal rates prior to 1986 (53 mortalities from 1972-1988), but have not
experienced high mortalities past 1986 (3 mortalities) and therefore they were not considered in the analysis.

A summary of changes in distributions of age as a function of sex was conducted for wildlife management areas that had suitable sample sizes for the analysis. Unlike the previous GBMA based analysis, this analysis considers periods prior to 1986 since much of the mortality and removals occurred for many of the areas during this time period (figure 14).


Figure 14: Changes in distribution of ages for male and female bear mortalities for selected wildlife management units. Midpoints correspond to pooled years as listed in Table 4.

A decline in age of females is potentially evident in wildlife management units 354 and 356 from 1991-2002; however, this may also be an artefact of sparse sample size of aged females. As with most areas, the number of mortalities declined in recent years therefore creating potential sparse sample size issues. In addition, the historic distribution of ages (1972-1986) suggests harvest of younger females, which may confound the results. No other trends are evident for other wildlife management units.

## Summary of harvest and hunter success data

Table 7 provides a summary of harvest and license data for each GBMA from 19882002. A harvest bear was any bear with an LH (legal harvest) code in the mortality database. All other form of mortality are not considered in this analysis.
Table 7 shows that the number of licenses allotted is small for some GBMA's. For example GBMA's 6 and 7 had only 3-4 licenses issued after 1998 in contrast to GBMA 3A and 2B, which had many more licenses, issued. Note that the number of license issued has decreased substantially for GBMA's 4A, 4C, and 6\&7.

Table 7: Summary of harvest and licenses allotted for GBMA's 1988-2002

| GBMA | Statistic | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | harvest | 3 | 1 | 6 | 2 | 4 | 3 | 1 | 0 | 4 | 2 | 4 | 0 | 2 | 2 | 2 |
|  | licenses | 50 | 51 | 72 | 37 | 37 | 36 | 23 | 24 | 32 | 32 | 37 | 31 | 3 | 17 | 30 |
| 2A | harvest | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 1 | 0 |
|  | licenses | 11 | 11 | 9 | 8 | 9 | 11 | 16 | 19 | 9 | 5 | 4 | 2 | 2 | 7 | 5 |
| 2B | harvest | 2 | 1 | 2 | 0 | 4 | 2 | 2 | 1 | 5 | 4 | 3 | 4 | 5 | 0 | 1 |
|  | licenses | 76 | 51 | 64 | 33 | 39 | 55 | 33 | 34 | 38 | 27 | 32 | 35 | 24 | 13 | 19 |
| 3A | harvest | 0 | 0 | 1 | 2 | 1 | 0 | 0 | 1 | 4 | 0 | 1 | 8 | 2 | 7 | 2 |
|  | licenses | 48 | 48 | 45 | 21 | 23 | 35 | 45 | 46 | 49 | 42 | 45 | 55 | 40 | 31 | 19 |
| 3B | harvest | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 1 | 0 | 2 | 0 | 0 | 1 | 0 | 1 |
|  | licenses | 16 | 17 | 24 | 12 | 12 | 24 | 12 | 10 | 10 | 19 | 11 | 17 | 21 | 25 | 27 |
| 4A | harvest | 0 | 0 | 2 | 0 | 2 | 6 | 1 | 2 | 2 | 0 | 1 | 1 | 3 | 0 | 1 |
|  | licenses | 31 | 19 | 24 | 19 | 20 | 31 | 9 | 8 | 8 | 8 | 6 | 7 | 7 | 5 | 8 |
| 4B | harvest | 0 | 2 | 2 | 3 | 0 | 2 | 0 | 2 | 2 | 0 | 1 | 1 | 2 | 3 | 3 |
|  | licenses | 21 | 18 | 26 | 17 | 18 | 13 | 10 | 20 | 13 | 9 | 9 | 12 | 9 | 16 | 15 |
| 4C | harvest | 0 | 1 | 3 | 1 | 5 | 1 | 0 | 1 | 0 | 1 | 1 | 0 | 1 | 1 | 3 |
|  | licenses | 26 | 13 | 15 | 16 | 15 | 6 | 4 | 3 | 3 | 4 | 3 | 3 | 4 | 4 | 3 |
| 6\&7 | harvest | 3 | 1 | 3 | 1 | 2 | 2 | 0 | 1 | 1 | 1 | 1 | 3 | 3 | 3 | 2 |
|  | licenses | 25 | 18 | 14 | 14 | 19 | 11 | 5 | 5 | 6 | 5 | 4 | 5 | 4 | 4 | 3 |

Hunter success (number bears harvested/number licenses allotted) was summarized to determine if there were any dominant trends that might indicate changes in population status. For example, a decreasing harvest success rate might indicate that population's are declining making it difficult to successfully hunt bears in a GBMA. Unfortunately, many other factors can influence hunter success such as changes in access to remote areas over time and hunter skill. The problem of changing access is a particularly difficult issue. For example, a population might be declining, however, increased access would still keep hunter success high. Additional information about hunter days spent for each year might bring more resolution to this index. Given these issue the results of this analysis should be viewed with extreme caution. See Miller (1989); and Garshelis (1993) for more cautions about the interpretation of harvest trends.

One other issue with hunter success data is the great degree of yearly variation, which makes it difficult to interpret raw data. For this reason, a smoothing spline routine was used to interpolate trend from data points. The spline data smoothing method can be conceptualized as a technique which undertakes non-linear regression analyses on small segments of data. It then connects the individual regression lines together in a method which optimizes the fit of each data point. This allows infinite flexibility in the shape of
relationships that splines can accommodate. The tension of the spline, which is a function of the number of data points used for each individual regression analysis, was varied iteratively to minimize noise or variance in the data set (Brown and Rothery 1993; Ellison 1993; SAS Institute 1997). The results of the spline analysis are shown along with the raw data points to allow further evaluation of trends in the data.
No apparent long-term trends are evident from this analysis. Success rate appears to be stable or increasing slightly in most GBMA' s. The actual number of licenses allotted can influence the interpretation of success rates. For example, the success rate for to be high in GBMA's 6 and 7 until 2000 when it decreases to O. However, the number of licenses allotted for the latter years of this analysis varies from 3 to 5 . The low sample size of licenses allotted basically makes the estimation of success rate for G BMA' s 6 and 7 problematic at best. A similar issue is evident with GBMA 4c in which only 3 licenses were issued with 3 kills in 2002 resulting in a success rate of 1 . To confront the issue we attempted to pool adjacent years, however, trends were still not evident from the data.

As mentioned earlier, data about hunter days spent would allow a better quantification of the degree of effort expended per hunter. This information might allow a better index of hunter success. In addition, changes in hunter access should be considered. For example, changes in kill locations over time might reveal if bears are being hunted in more remote areas than previous years.


Figure 15: Trends in harvest success by GBMA. A spline smoothing routine was used to evaluate trend in the data. Raw data points are also shown.

## Problem Bear Management and Relocations

As part of an overall assessment of provincial grizzly bear mortalities it is important to recognize and appreciate that problem bear management in Alberta can, and will, playa key role in the long-term conservation of this species. Figure 16 shows overall problem bear management actions (captures and relocations) for the period 1974-1997. Although UTM's for the period 1998-2001 were not available, the same general spatial trend was apparent in these years. This map clearly shows common geographic areas of problem bear occurrences and also identifies common areas where the department has been relocating these problem bears. There has been little or no change in the location or reasons for problem bear actions in these areas since the publication of the 1990 Alberta Grizzly Bear Management Plan. This suggests that our current management approach to bear problems in this area appears to be inadequate and that new approaches or efforts are required.


Figure 16. Spatial distribution of problem bear captures and relocations.

Historically problem bears that were relocated were treated as a loss to the BMA where they were captured and added to the BMA of relocation. This practise served the general requirements of the previous allocation formula but did not take into account either; a). the mortality rates of relocated bears (Blanchard and Knight 1996)( estimated to be at least 30\%), or b). what impact the release of a large number of bears would have on the resident bear population. At this time biologists do not have an understanding of this latter issue.

From a population management standpoint it is important to reduce the number of grizzly bears that are being relocated in the province to increase the overall survival rate of this cohort within the population. An increased understanding of the survival and population impacts of problem bear relocations would be extremely valuable.

Effective translocation management will require a more rigorous assessment of habitat suitability and human-use factors of relocation sites to avoid placing relocated bears in areas where the likelihood of human caused mortality is high. The last provincial assessment of problem bear relocation sites appears to have been made in 1982 when land-use patterns were significantly different than today. Given the geographical concentration of problem bear captures it would also be prudent to have department staff develop programs and approaches to help reduce the occurrence of problem bears by working with stakeholders to identify and reduce attractants, etc.

## Other Considerations

Grizzly bear hunting in Alberta is strictly controlled through a regulated draw and quota system. The number of permits issued is based on the most recent annual population estimate and by following an approach of sustained yield. Hunting is allowed between April 1 and May 31st each spring, and protection is extended to all grizzly bears part of a group of two or more. In addition our regulations prohibit the harvest of grizzlies less than two years of age and females accompanied by cubs. Hunters are also prohibited from using bait when hunting grizzly bears.

In recent times, in areas where grizzly bear research activities have been underway, department staff has found that some hunters are hunting in and around active trapping sites. Since some research activities require the capture and collaring of bears and bait is used for this purpose it appears that some hunters are benefiting from hunting near this bait. In fact some hunters appear to have entered closed areas (under Section 6.1 of the Wildlife Act) where active trap sites are present in pursuit of a grizzly bear.

As a result of this conflict, consideration should be given to closing identified BMA's where active grizzly bear trapping efforts (using baits) for research purposes, are underway.

Jurisdictions where black bear hunting takes place in habitats shared with grizzly bears have realized that this can lead to an increase in the mortality of grizzly bears. At times both black and grizzly bears can be misidentified and an illegal harvest can occur.

Alberta also faces similar challenges in having a fall black and spring black bears where grizzly bears are sometimes killed due to misidentification of this species. There have been two common approaches used in other jurisdictions to deal with this issue. First management decisions have been made to prohibit black bear hunting in management units where both black bears and grizzly bears are found. The second approach is to require all bear hunters to take and pass a formal test in which they have to correctly identify the species of bear shown in a series of photographs as recently implemented in Montana. With this approach management agencies also have the ability to review with hunters the requirements of harvest registration and review information on the collection of biological samples. Alberta would benefit from the implementation of this type of system in an effort to reduce the number of "misidentified" grizzly bear mortalities.

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Appendix A


Appendix B


Appendix C.

